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Title

Life Cycle Assessment to compare the environmental impacts of different wheat production systems

Authors

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Life Cycle Assessment to compare the environmental impacts of different wheat production systems

Abstract

The share of fossil-based resources in the implementation of agricultural activities in Iran is very high. In this context, it is important to determine the environmental impacts of energy use in agricultural activities. The purpose of this study was to evaluate and compare the energy consumption and total environmental impacts of irrigated and rainfed wheat production in central Iran, Mahyar plain. For this purpose, data were collected from 120 irrigated and 90 rainfed wheat farms in three different farm size (<2 ha, 2-4 ha and more than 4 ha), through questionnaires and site visits. In this study, standard ISO life cycle assessment methodology was used to evaluate the total impact of all consuming inputs on environmental pollution and show the main hotspot in the production chain. Results of energy analyses showed that large farms used more energy per unit of farm land than that of small farms. Farmers also used higher input energy for irrigated wheat in comparison with rainfed system. The overall energy use efficiency for per unit farm of irrigated wheat production was only half of rainfed wheat production. Results of Life Cycle Assessment (LCA) models show that rainfed wheat actually produced more pollutions than irrigated wheat production because of lower yield in ha. In this study, the rate of Abiotic Depletion (AD) and Acidification (AC) impact were 0.002-0.003 kg Sb eq and 8.991-11.863 kg SO₂ eq for wheat production (irrigated and rainfed), respectively. Also the Ozone Layer Depletion (OLD) and Photochemical Oxidation (PO) were calculated 0.00002-0.00004 kg CFC11 eq and 0.145-0.174 kg C₂H₄ eq for wheat production (irrigated and rainfed), respectively. The results showed that the main hotspots for irrigated and rainfed wheat production were chemical fertilizers and diesel fuel. Based on the results, it is suggested to use more intensive cropping systems (such as solar greenhouse) to decrease the intensity of input energy and increase the output level of productions with minimum environmental pollutions.

Keywords

Sustainable agriculture, wheat production, energy consumption, environmental analysis

Nomenclature

E_o	Energy output (GJ/ha)	E_f	Fertilizer energy (MJ/ha)
E_i	Energy input (GJ/ha)	W_f	Weight of fertilizer (kg/ha)
E_p	Energy productivity (kg/GJ)	E_k	Existing energy in fertilizer (MJ/kg)
O_p	Output production (kg/ha)	F_c	Fuel consumption required for agricultural practice (L/ha)
E_e	Electricity energy (MJ/ha)	F_{hr}	Fuel required (L/hr)
N_E	Net energy (GJ/ha)	T	working time of machinery (hr/ha)
M_E	Machinery energy (GJ/ha)	E_m	Energy provided by a machine (coefficient equal to 62.7 MJ kg ⁻¹)
W	weight of machine (kg)	E_t	Effective lifetime of tractor (hr)
Q_h	Total machine working hours during an agricultural season	ρ	Water density (kg/m ³)
Φ	Water flow rate (m ³ /ha)	g	gravitational acceleration (m/s ²)
E_R	Energy Ration (-)	H	Total dynamic well head (m)
ε_2	Efficiency of energy and power (0.18 -0.22 for electro pump and 0.25-0.30 for diesel)	ε_1	Pumping efficiency varying between 0.7 and 0.9
TE	Terrestrial Ecotoxicity (g 1,4-DB eq)	AD	Abiotic Depletion (g Sb eq)
LCA	Life Cycle Assessment	AC	Acidification (g SO ₂ eq)
LCI	Life Cycle Inventory	EP	Eutrophication (g SO ₂ eq)
$LCIA$	Life Cycle Impact Assessment	G	Global Warming (kg CO ₂ eq)
FU	Functional Unit	HT	Human Toxicity (g 1,4-DB eq)
OLD	Ozone Layer Depletion (g CFC-11 eq)	ME	Marine aquatic Ecotoxicity (g 1,4-DB eq)
FE	Fresh water aquatic Ecotoxicity (g 1,4-DB eq)	PO	Photochemical Oxidation (g C ₂ H ₄ eq)

1.Introduction

Wheat is one of the important cereals in the world and has a major rank among low-income household in the world (Taki et al, 2016). Planted area of wheat was 4.09 million ha in 2015-2016 in the world. Total harvested cereals in 2015-2016 were 18 million tons which the share of wheat was about 63.16%, followed by barely (17.55%), paddy (12.87%) and corn (6.40%) (Taki et al, 2016).

One of the key criteria for choosing strategic approaches to achieving sustainable agriculture is to increasing the efficiency of production (Bundschuh, and Chen, 2014). Efficiency can be increased in agricultural production by more sustainable use of resources, such as labor, chemical fertilizers, water and fuel. Also application of new technologies and attention to life cycles of byproduct processing can increase the agricultural efficiency (Maraseni, et al., 2015).

Currently, in most developing countries, a lot of fossil energy sources are using to produce agricultural productions, which often lead to air, soil and water pollution (Nemecek et al, 2011). This high energy consumption can loss the valuable fossil fuels and create great energy challenges for future generations (Khoshnevisan et al, 2013a).

LCA as a powerful method was used to investigate the environmental impacts of energy sources in many researches (Taki et al, 2016). This method can determine the total air, water and soil pollution during production a particular product (Chen at al., 2010). A lot of researchers used this method for analyzing the energy consumption of agricultural products. Some of them were focused on cereal production. For example, Taghavifare et al (2015) and Khoshnevisan et al (2015) focused on Iran wheat production. Brentrup et al (2004a and b) analyzed the effect of

chemical fertilizers on crop yield in the UK. Achten and Van Acker (2016) reviewed the wheat production systems in Europe. Kim et al (2009) evaluated LCA of corn grain and corn stover in the United States. Biswas et al (2008) presented a LCA analysis of wheat production in south-western Australia.

Iran is dependent on excessive fossil-based energy resources within agricultural activities and CO₂ emissions because of agricultural activities are high. This study will indicate the environmental impacts of fossil based energy sources in the wheat production process in specific region of Iran. However, the results of environmental impacts of fossil based energy sources in wheat production will be compared to those of previous LCA studies.

In this study, ISO LCA method (ISO, 2006) was applied for production of irrigated and rainfed wheat in an Iranian farmer cooperative, located in the center of Iran, based on the first phases of life cycle (i.e. from tillage to harvest), with the aim to evaluate their energy flow, environmental performance and the hotspots in the production chains. This area is very important for wheat production in Isfahan province. Considering the mechanization development plans and the necessity of integrating the farms in Iran and also growing concerns about environmental and economic issues, the present study aimed to investigate the relation between farm size and environmental impacts, identify hotspots in the production chains and evaluate energy flow. Actually, it is worthy to be known that, how farm size affects the environmental impacts. So, in the present study, farm size was investigated from environmental point of view in three categories. Also considering the lack of freshwater resources and efforts to find the deeper wells, present study makes an accurate assessment of energy amount for wheat production in this plain and provide solutions to this problem. Some recent researches evaluated the energy consumption in Isfahan province (Khoshnevisan et al, 2013a,b,c) but in this study only Mahyar plain, located in Shahreza city will be evaluated. So the purpose of this study is to accurately assess the amount of energy and its pollution effects for wheat production in this plain and finally provide some solutions for this problem.

2. Methodology

2.1. Case study region

Shahreza is located 508 km south to Tehran and about 80 km south west of Isfahan. Zard Kooh mountain range runs from northwest to southeast of this city, enjoying a cold climate (Fig.1). In this study, two types of wheat farms were evaluated: irrigated and rainfed. For each category, three samples were selected, i.e small (<2 ha), medium (2-4 ha) and large farms (> 4 ha). The data were collected from 120 irrigated and 90 rainfed wheat farms in Shahreza city in Isfahan province using face to face questionnaire method.

In this research, experimental database was randomly selected from the rural communities in the research region. For calculating the sample size, Neyman technique was used (Taki et al, 2013):

$$n = \frac{\sum N_s P_s}{N^2 D^2 + \sum N_s P_s^2} \quad (1)$$

where n is the sample size of total data; N is the number of whole population; N_s is the number of the population in the s stratification; P_s is the standard deviation in the s stratification p_s^2 , is the variance in the s stratification, D² is equal to $\frac{d^2}{z^2}$; d is the precision and z is the reliability coefficient (1.96, which represents 95% reliability). Thus the total sample number was 120

irrigated and 90 rainfed wheat farms. Table 1 shows the summary of the sampling wheat farms, according to the size and type of them.

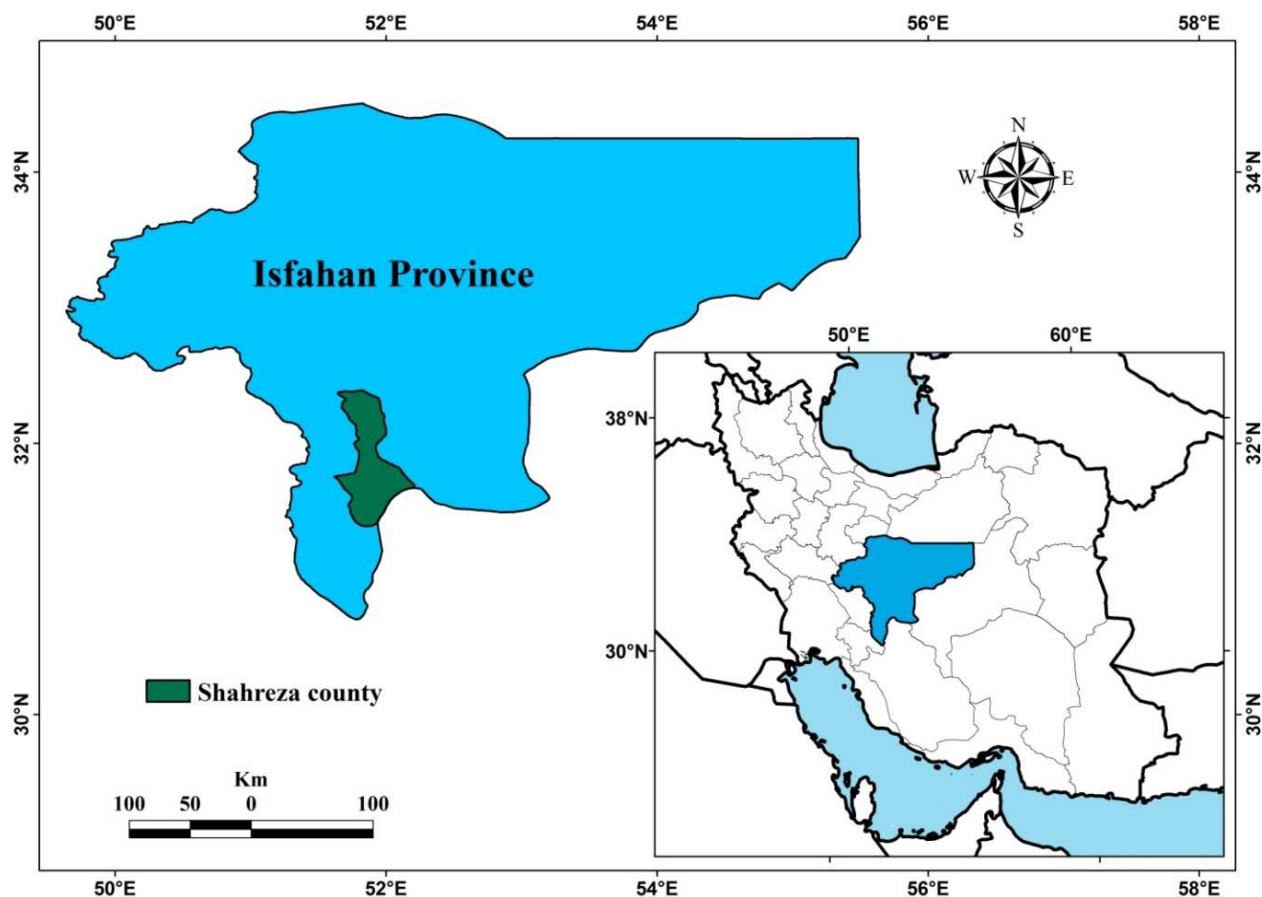


Fig.1. Location of the studied area.

Table.1. Summary of wheat farm types and number of sampling for each one

Type of wheat farms	(< 2 ha)	(2-4 ha)	(> 4 ha)
Irrigated wheat farms	50	40	30
Rainfed wheat farms	20	40	30
Total	70	80	60

2.2. Energy analysis method

In this research, total energy using in wheat production was analyzed (Natural sources of energy such as solar radiation, are not considered). In this study, energy consumption in total process of wheat production (human, diesel fuel, chemical and farmyard fertilizers, electricity, machinery, pesticides and seed) was calculated based interviews and oral questions (Chen at al., 2010). Also, the output energy was calculated only for grain production because it is the main commodity produced. For converting all the materials into the energy equivalents, Table 2 was used.

Table.2. Energy coefficients of different materials for wheat production.

Inputs	Unit	Energy equivalent (MJ unit ⁻¹)	Reference
1.Human labor	Man	h	Ranjbar et al, 2013
	Woman	h	
2.Chemical fertilizer and farmyard manure	N	kg	Taki et al, 2012c
	P ₂ O ₅	kg	
	K ₂ O	kg	
	farmyard manure	kg	
3.Chemical poison	Pesticides	kg	Ranjbar et al, 2013
	Herbicide	kg	
4.Agricultural machinery	Tractor	kg year	Abdi et al, 2012
	Implement and machinery	kg year	
	Combine harvester and mower	kg year	
	Improved seed or hybrid	kg	
5. Seed	General seed	kg	Abdi et al, 2013
6.Diesel fuel	L	47.80	Ranjbar et al, 2013
7.Electricity	kWh	3.60	Taki et al, 2012b

It is clear that energy conversion factors may be different for different regions, but for a particular area, these factors can be used to let the researchers to have a good comparison between using of energy for production a specific product. In this study, total energy using for wheat production can calculate to below categories:

2.2.1. Fuel consumption

Time duration for each section of planting was counted from starting of each agricultural action and also applying the information from the expert farmers. In this study, diesel fuel consumption was calculated using (Taghavifar and Mardani, 2015):

$$F_c = F_{hr} \times T \quad (2)$$

2.2.2. Electricity

In wheat production, electricity is consumed for irrigation. Water for irrigation in this region was mostly pumped from deep well by huge electric pumps. One of the serious problems in this region is the lack of adequate water because of excessive consumption and using old irrigation systems. So, one of the largest energy consuming sources in wheat production in this region is the energy needed for water pumping by electric pumps. (Taki et al, 2016):

$$E_e = \frac{\rho \times g \times H \times \Phi}{\varepsilon_1 \times \varepsilon_2} \quad (3)$$

2.2.3. Human Labor

Generally, human power is considered as the muscular strength of a working man (Nawi et al., 2012). Multiplying the number of workers with the energy equivalent of 1 hour of man labor and the total working time by the workers can calculate the human power for all agricultural activities. (Nabavi-Pelesaraei et al, 2013).

2.2.4. Embodied energy of machinery

For calculating the energy consumption and share of each machine in wheat production, machinery effective lifetime, average area and weight are important factors. The following equation was used to calculate the share and energy of each machine and implement (Ramedani et al, 2011):

$$M_E = E_m \times \frac{W}{E_t} \times Q_h \quad (4)$$

2.2.5. Fertilizer

Energy for fertilizer was calculated based on N, P₂O₅ and K₂O energy equivalent by the following equation (Taghavifar and Mardani, 2015):

$$E_f = W_f \times E_k \quad (5)$$

2.2.6. Spraying

For calculating the energy equivalent of pesticides used in agricultural production, it is possible to use the multiplication of the amount of active ingredient of each poison (kg/L) in the equivalent energy used (Table 2) for its production (Taghavifar and Mardani, 2015).

2.2.7. Energy indicators

Some energy indicators can be used to calculate the energy efficiency of agricultural production. (Yuan and Peng; 2017; Taki et al, 2018). Based on the energy equivalents (Table 2), energy use efficiency (energy ratio), energy productivity and net energy gain were computed as (Taki et al, 2012c):

$$E_R = \frac{E_o}{E_i} \quad (6)$$

$$E_P = \frac{O_p}{E_i} \quad (7)$$

$$N_E = E_o - E_i \quad (8)$$

2.3. LCA methodology

LCA is a widely used method for assessing and studying the performance of crop production (Liamsangan and Gheewala, 2008). LCA was defined as an assessment and evaluation of the outputs, inputs and the total environmental impacts of a product system during its life cycle (ISO 14040, 2006) (Nabavi-Pelesaraei et al, 2017). The amicability of LCA in numerous previous studies was evaluated (Cleary, 2009). Based on the ISO 14040, the structure of LCA includes four separate stages that contribute to a unified approach (ISO, 2006): 1. Goal and area definition. 2. Life cycle inventory (LCI). 3. Life cycle impact assessment (LCIA). 4. Interpretation (Nabavi-Pelesaraei et al, 2017).

2.3.1. Structure of LCA

To define the aim and scope of the study, the suggestions based on the International Reference Life Cycle Data System (IRLCDS) Handbook (European Commission, 2010) was applied. This research is a cradle to gate study and accordingly, the objective of this study is to assess the environmental performance of irrigated and rainfed wheat production. LCA present and analysis total environmental impact of energy consumption based on the system boundaries and Functional Unit (FU) (Rebitzer et al., 2004). FU is usually defined as the system's output. System boundary of this study covers all farm operations for wheat production. Fig.2 shows the system boundary for irrigated wheat production (Rainfed system boundary is very similar to irrigated but has not water for irrigation). As it can be seen, for irrigated wheat, electricity is used for pumping water and therefore the only difference between rainfed and irrigated wheat processing is electricity (Pahlavan et al, 2011). Direct emissions are the result of the operations but the source of indirect emissions is the process of procuring diesel fuels, electricity (produced and consumed) and other products. The FU adopted was the production of 1 ton of wheat during a production period.

System boundary for irrigated and rainfed wheat production

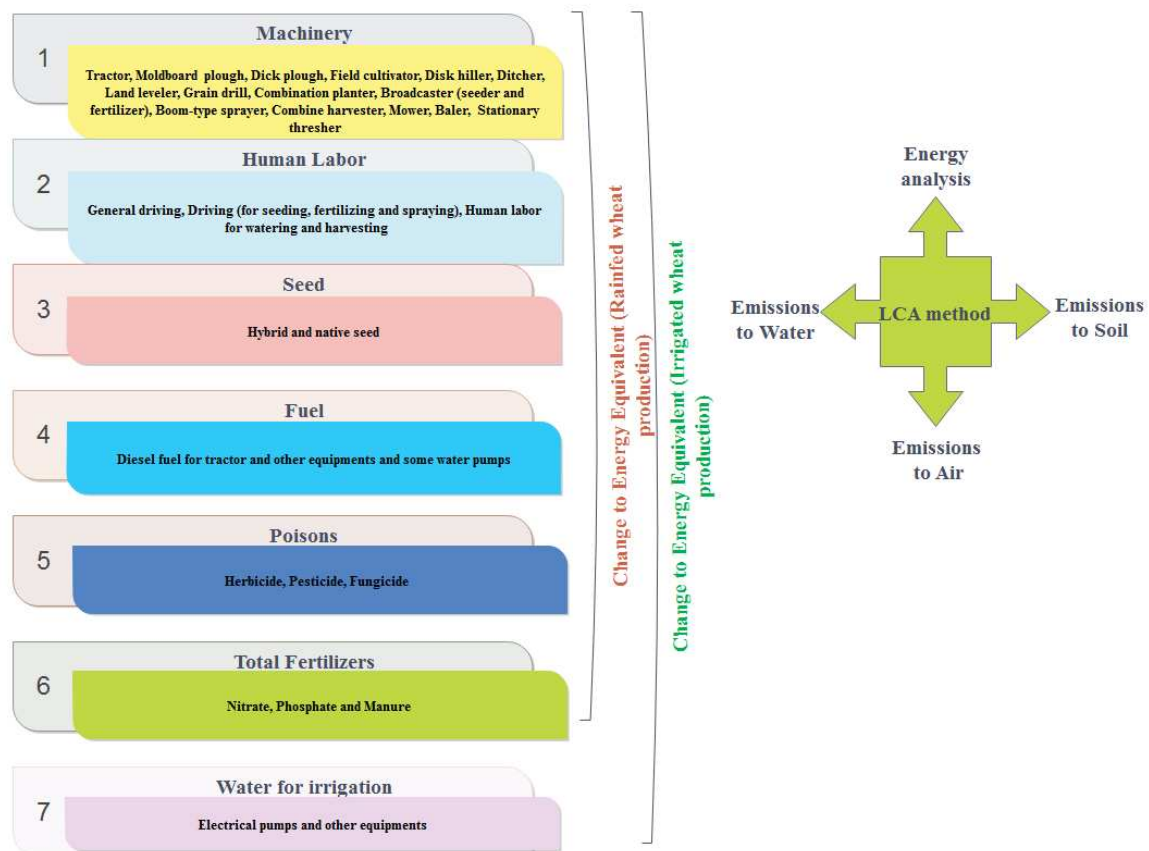


Fig.2. System boundary of LCA process for irrigated and rainfed wheat production.**2.3.2. Life Cycle Inventory (LCI) Analysis**

In the second phase of LCA method, the product system is defined (Blanco et al, 2015). In this stage, total inputs and outputs can change to quantify situation (Kouchaki-Penchah et al, 2017). In this section, LCI data for were acquired (1 ton of wheat). Inventory data in the foreground system (actual crop production) consisted of actual farm practices obtained through face-to-face questionnaire method with wheat farmers (Table 1) (Rivela et al, 2006). Other inventory data such as machinery, nitrogen, phosphate, farmyard manure and biocides were obtained from literature and systematic software databases (Ecoinvent 3.0 and Agri-footprint) (Wernet et al., 2016). Emission of diesel fuel was added from Nemecek and Kagi (2007) and Nielsen et al (2005) (Table 3).

Table.3. Life cycle inventory data for a MJ traction in Ecoinvent database.

Emissions	Amount (g/MJ diesel)
CO ₂	74.5
SO ₂	2.41E-02
CH ₄	3.08E-03
Benzene	2.39E-07
Cd	1.19E-06
Cr	1.19E-06
(Cu	4.06E-05
N ₂ O	2.86E-03
Ni	1.67E-06
Zn	2.39E-05
Benzo (a) pyrene	7.16E-07
NH ₃	4.77E-04
Se	2.39E-07
PAH (polycyclic hydrocarbons)	7.85E-05
Hydro carbons (HC, as NMVOC)	6.80E-02
NO _x	1.06
CO	1.50E-01
Particulates (b2.5 µm)	1.07E-01

2.3.3. Life Cycle Impact Assessment (LCIA)

The aim of the LCIA, is to evaluate the amount and effect of the potential environmental impacts of a product system (Nabavi-Pelesaraei et al, 2014). The impact categories were analyzed based on CML-IA baseline V3.01/EU25 (Guinee, 2002) method and include: human toxicity (HT), acidification (AC), eutrophication (EP), ozone layer depletion (OLD), fresh water aquatic ecotoxicity (FE), abiotic depletion (AD), global warming (GW), terrestrial ecotoxicity (TE), marine aquatic ecotoxicity (ME) and photochemical oxidation (PO) (Durlinger et al., 2014). All these impact categories can affect ecosystem quality, human health, climate change and resources as damage categories. Fig.3 shows the links between impact and damage categories. For energy analysis and evaluate the potential environmental impact, Excel 2013 spreadsheets and SimaPro V8.0 software were used.

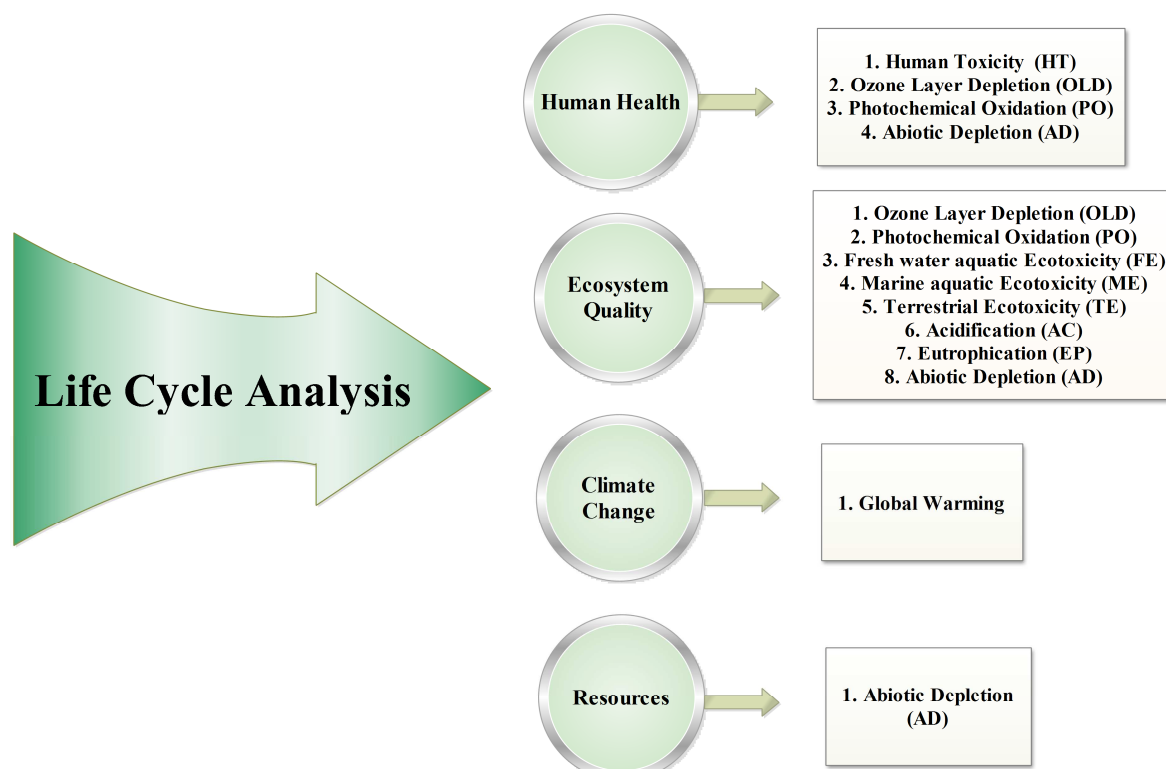


Fig.3. Links between impact categories and damage categories

3. Results and discussion

3.1. Energy consumption

Table 4 shows a summary of energy consumption (GJ ha^{-1}) for wheat production in different farm sizes. Total energy consumption in wheat production (irrigated and rainfed) and output energy were 23.41, 9.35 and 64.84, 35.22 GJ/ha , respectively. As it can be seen, in all farm types, diesel fuel had the highest impact on energy output except for small rainfed farms (<2 ha-rainfed). In small farms, total chemical fertilizer and poisons had the highest impact on output energy (The farmers used to spread the fertilizer with hand and this method can increase the fertilizer consumption). Khoshnevisan et al (2013a) reported that total input energy for wheat production in Isfahan province were 80.40, 79.29 and 81.11 GJ/ha for small (<1 ha), medium (1-3 ha) and large (>3 ha) farms, respectively. The results in this research are significantly higher than present results. In that research, electricity and chemical fertilizer had the most important effects on output energy (49% and 29%, respectively).

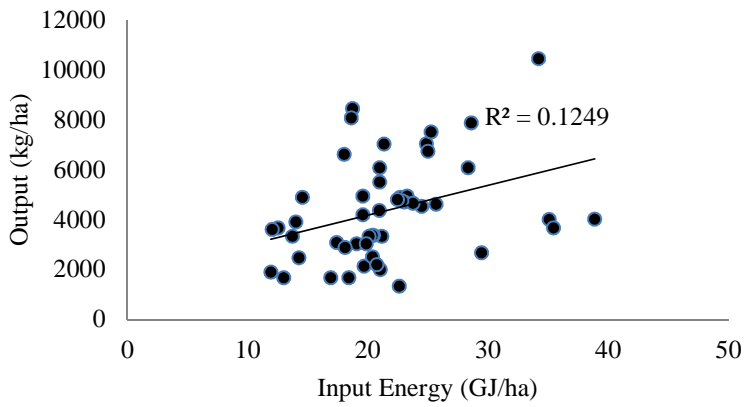
In a similar study, Safa et al (1999) reported that total input energy for irrigated and rainfed (dryland) wheat production in New Zealand were 25.6 and 17.45 GJ/ha , respectively. In that research, chemical fertilizer and electricity had the highest share in irrigated farms. In the present research, some of farmers used diesel fuel engines for water pumping. This type of energy was calculated for diesel fuels. In this region, because of increasing of drought period and also existing of deep wells, farmers had to use much energy for water extraction. Similar results were reported by Abdollahpour and Zaree et al (2009). The authors showed that total input and output energy for irrigated wheat production was 25.67 and 21.01 GJ/ha at west of Iran (Kermanshah province).

The results of present study showed that in large farms (irrigated), the total input energy used per hectare was higher than small farms, but the total output energy per hectare in small farms was actually higher than large farms. This results showed that using of technology was not effective in irrigated wheat farms and will be worse by increasing in farm sizing. Actually, after the White Revolution in 1963-1978, farm lands were divided between farmers and so their sizes became smaller than in the past. The results of Table 4 showed that, total input of machinery will increase by changing the farm sizing (small to large).

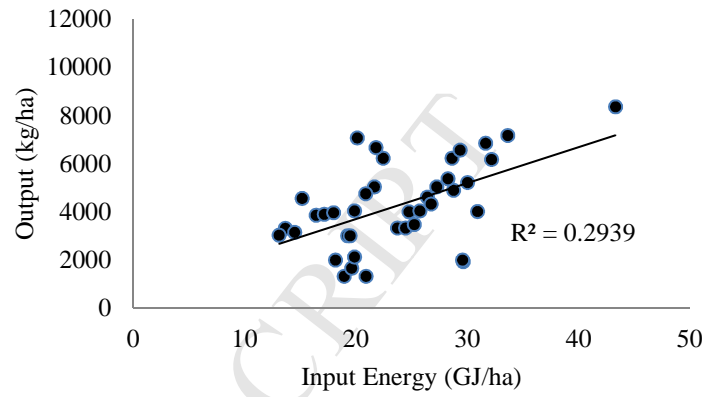
Table.4. Energy inputs and output for wheat production (GJ/ha).

Item	Irrigated wheat farms						Rainfed wheat farms					
	Small (< 2 ha)		Medium (2-4 ha)		Large (> 4 ha)		Small (< 2 ha)		Medium (2-4 ha)		Large (> 4 ha)	
	Mean	Std	Mean	Std	Mean	Std	Mean	Std	Mean	Std	Mean	Std
Machinery	0.66	0.20	0.84	0.24	0.83	0.20	0.31	0.08	0.41	0.16	0.49	0.17
Water for irrigation	3.19	3.51	4.54	4.53	5.46	4.68	0	0	0	0	0	0
Human labor	0.20	0.11	0.19	0.09	0.13	0.08	0.14	0.08	0.10	0.07	0.05	0.05
Diesel fuel	6.52	1.53	7.64	2.09	7.22	1.74	3.69	0.56	4.30	1.11	4.44	1.02
Total chemical Fertilizer	5.54	0.87	5.21	0.96	6.34	0.92	3.77	0.84	3.18	0.83	2.85	0.67
Pesticides	0.10	0.10	0.44	0.13	0.46	0.09	0.08	0.10	0.09	0.10	0.09	0.06
Herbicides	0.37	0.12	0.16	0.08	0.15	0.09	0.21	0.24	0.25	0.24	0.27	0.24
Fungicides	0.04	0.05	0.08	0.05	0.06	0.05	0.06	0.03	0.05	0.04	0.05	0.04
Seed	4.86	3.02	4.68	3.16	4.33	2.66	1.08	0.84	1.10	0.32	1.00	0.56
Total input energy	21.48	-	23.78	-	24.98	-	9.34	-	9.47	-	9.24	-
Average of input energy (irrigated and rainfed)			23.41						9.35			
Output energy (Grain)	59.80	-	62.60	-	72.11	-	38.48	-	35.20	-	32.00	-
Average of output energy (for irrigated and rainfed)			64.84						35.22			

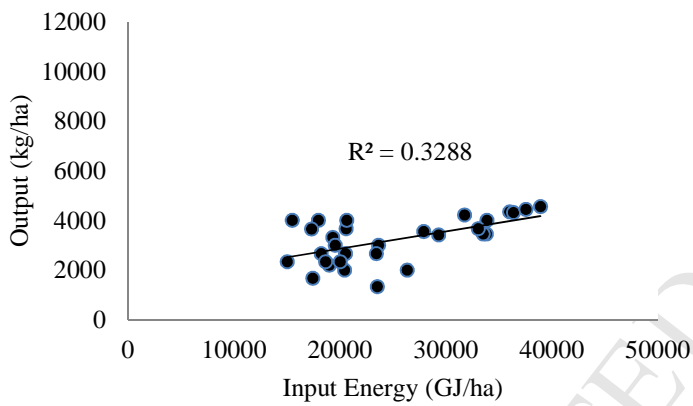
Fig.4 shows the correlation between input energy and yield for irrigated and rainfed wheat production. In this research, for irrigated wheat, diesel fuel has the highest share in total inputs (29%) followed by chemical fertilizers (25%). Diesel fuel consumed the most energy of total energy inputs in rainfed farms (48%), followed by total fertilizers and poisons (35%). The share of total fertilizers in small rainfed farms was higher than diesel fuel, because in these farms, humans usually spread the fertilizers and the farmers did not use machinery for this operation. Ghorbani et al. (2011) reported that total energy input for irrigated and rainfed wheat production systems in Northern Khorasan province of Iran were 45.37 and 9.35 GJ/ha, respectively.



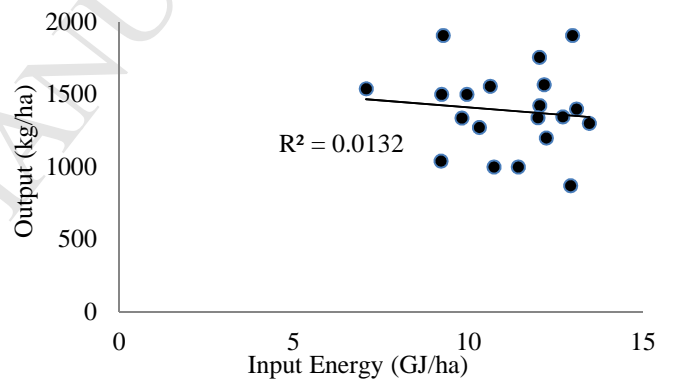
< 2 ha-Irrigated



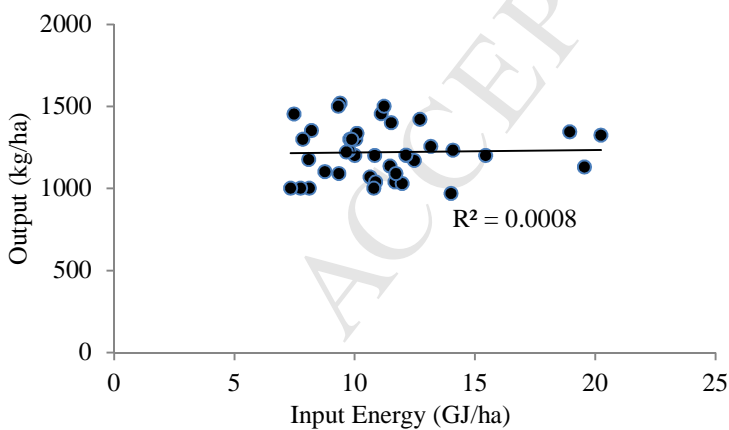
2-4 ha-Irrigated



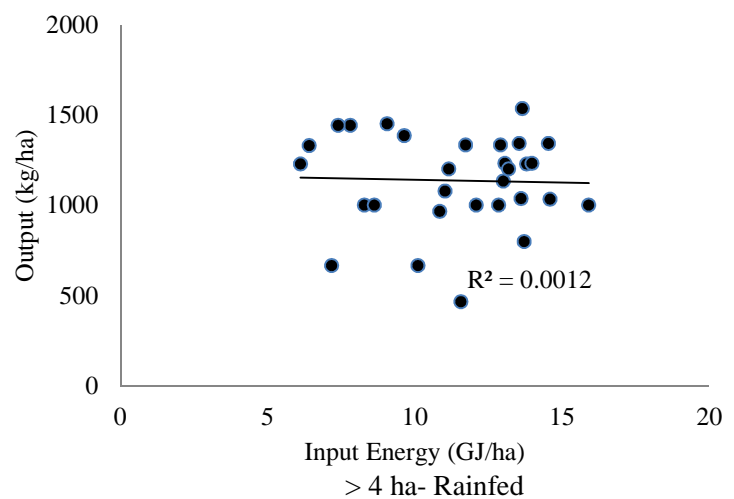
> 4 ha-Irrigated



< 2 ha-Rainfed



2-4 ha- Rainfed



> 4 ha- Rainfed

Fig.4. Correlation between energy consumption (GJ/ha) and yield (kg/ha) in irrigated and rainfed wheat production

The results indicated that chemical fertilizers, electricity and diesel fuel inputs were 10.95, 16.84 and 4.32 GJ/ha respectively in irrigated wheat farms. Similar results were reported in other studies for different crops such as irrigated and dryland chickpea production systems (Koocheki et al., 2011) and sugar beet (Asgharipour et al., 2012).

Positive correlation between yield and energy consumption for irrigated and negative correlation for rainfed wheat production farms was the most attractive finding of this research. This indicate can able the farmers to increase energy input and receive more yield at irrigated farms. The reason of this fact is mostly because of irrigation and fertilizer, that prepares better situation for crop growing. However, finding a balanced solution that could optimize the constraints on production, environmental and financial benefits simultaneously is one of the most important purposes in sustainable farming (Safa et al, 2011). As it can be seen, for irrigated farms (especially for large lands), the correlation between input energy and output yield was higher than other (0.32). The energy ratio, energy productivity and net energy of irrigated and rainfed wheat production are tabulated in Table 5.

Table.5. The share of each category on total energy inputs for irrigated and rainfed wheat production.

Item	Irrigated wheat farms			Rainfed wheat farms				
	Unit	(< 2 ha)	(2-4 ha)	(> 4 ha)	Unit	(< 2 ha)	(2-4 ha)	(> 4 ha)
Energy ratio	-	2.78	2.63	2.89	-	4.11	3.71	3.46
Net energy	GJ/ha	38.32	38.83	47.14	GJ/ha	29.14	25.73	22.76
Energy productivity	kg/GJ	145.49	122.50	129.90	kg/GJ	149.60	127.45	123.45
^a Direct energy	GJ/ha	9.91 (46%)	12.37 (52%)	12.81 (51%)	GJ/ha	3.83 (41%)	4.41 (46%)	4.49 (48%)
^b Indirect energy	GJ/ha	11.57	11.40	12.17	GJ/ha	5.51	5.07	4.74
^c Renewable energy	GJ/ha	5.06	4.87	4.46	GJ/ha	1.22	1.20	1.05
^d Non-renewable energy	GJ/ha	16.42	18.90	20.51	GJ/ha	5.51	8.25	8.19
Total energy	GJ/ha	21.48	23.77	24.97	GJ/ha	9.34	9.47	9.24

^a Includes diesel fuel, human labor and water for irrigation

^b Includes machinery, farmyard manure, biocide, seed and chemical fertilizer

^c Includes farmyard manure, seed and human labor

^d Includes chemical fertilizer, diesel fuel, machinery and water for irrigation

In total energy researches, energy ratio is a general index to evaluate the efficiency of sources. (Kuesters and Lammell, 1999). In this research, this index for different land sizes of irrigated and rainfed wheat production was 2.60-2.90 and 3.10-4.10, respectively. The average of energy use efficiency in rainfed wheat production was nearly more than twice higher than irrigated wheat production, which showed better efficiency in rainfed wheat production based on 1 ha wheat crop. the researches, evaluated the energy ratio for different crops. For example, it was found this indicate was 3.40 for rainfed and 1.4 for irrigated wheat production system (Ghorbani et al., 2011), 1.20 for irrigated and 2.90 for dryland chickpea production system (Koocheki et al., 2011), 13.40 for sugar beet (Asgharipour et al., 2012) and also 3 and 3.85 for irrigated and rainfed wheat production (Mondany et al, 2017). The results indicated that in irrigated wheat production, the highest and the lowest energy use efficiency were 2.63 and 2.89 for medium and

large farms, respectively. It shows that larger irrigated farms can produce wheat more effectively than smaller irrigated farms.

The results of this study showed that average energy productivity in irrigated and rainfed wheat production were 132.63 and 133.50 kg GJ⁻¹, respectively. Energy productivity is a very important indicator for evaluating the crop production systems and energy output. Energy productivity indicates sustainability and security in agricultural production systems. Several researchers have evaluated this index for agricultural crops production (Rezvani-Moghaddam et al., 2011; Koocheki et al., 2011; Ghorbani et al., 2011). In this study, net energy index shows consuming lots of extra energy in rainfed wheat. It demonstrated that in this region, the farmers have a very significant potential in saving energy. For increasing the energy efficiency systems, the farmers would often need governmental assistance and low-interest loans.

Fig.5 shows the share of two main types of chemical fertilizers in irrigated and rainfed wheat production. As it can be seen, in irrigated farms, chemical fertilizers were consumed more than in rainfed farms. The high final price of production, environmental pollutions and some human hazard risks are the consequences of this excessive consumption.

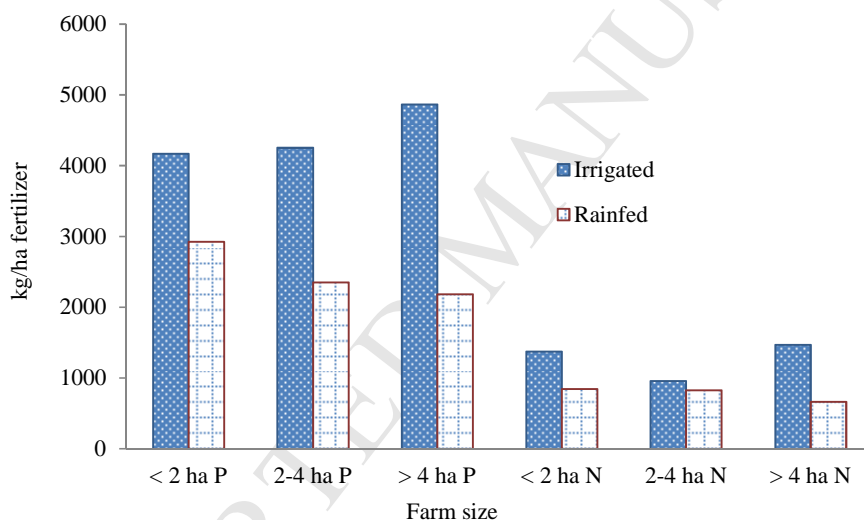


Fig.5. The share of two types of chemical fertilizer (N and P₂O₅) for irrigated and rainfed wheat production.

The share of direct, indirect, renewable, and non-renewable forms of energy is shown in Table 5. As it can be seen, the share of indirect energy in small and medium farms is higher than in large farms. In large farms, the application of machinery was higher than others. In this region, most of the farmers used hybrid seeds, so the equivalent energy of seed was higher than other researches. The share of direct energy in irrigated wheat production was higher than similar results for rainfed wheat production in all categories (small, medium and large farms) (Table 5). So, it seems that rainfed wheat production was more sustainable production system compared to irrigated wheat system. The results in Table 5 indicate that the share of renewable energy including seed, human labor and farmyard manure in irrigated was higher than rainfed wheat production. In all categories, the average of renewable energy for irrigated and rainfed wheat production is 4.80 and 1.16 GJ/ha, respectively. Fig. 6 shows the share of non-renewable and renewable energy in all categories for irrigated and rainfed wheat production. The excessive consumption of diesel fuel and electricity was the main reason for higher levels of non-

renewable energy in all agricultural productions. Mohammadi et al. (2014) reported that total input energy in wheat farms of Golestan province was 26.20 GJ/ha, which the share of direct and indirect energies was 58.8% and 41.2%, respectively. The share of renewable and non-renewable energies was 17.9% and 82.1%, respectively.

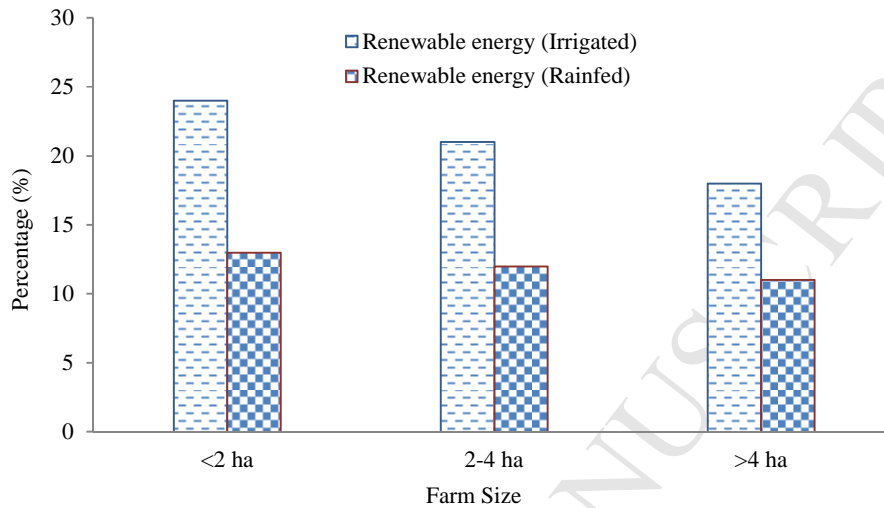


Fig.6. The share of renewable energy for irrigated and rained wheat production in different farm sizes.

3.2.LCA results

A brief summary of environmental impacts based on wheat production (per ton wheat in all periods of cultivation) for irrigated and rainfed wheat production is shown in Table 6.

Table.6. Values of the potential environmental impact of irrigated and rainfed wheat production per ton of wheat produced.

Impact Category	Unit	Irrigated Wheat	Rainfed Wheat
AD	kg Sb eq	0.002	0.003
GW	kg CO ₂ eq	317.81	380.16
OLD	kg CFC-11 eq	0.00002	0.00004
HT	kg 1,4-DCB eq	81.45	99.96
FE	kg 1,4-DCB eq	40.95	53.42
ME	kg 1,4-DCB eq	203,362.64	275,721.25
TE	kg 1,4-DCB eq	0.46	0.69
PO	kg C ₂ H ₄ eq	0.15	0.17
AC	kg SO ₂ eq	8.99	11.86
EP	kg PO ₄ ³⁻ eq	2.23	3.18

Primary LCA models show that rainfed wheat produced more pollutions than irrigated wheat system because of significantly lower yield. The reason of this fact is more energy consumption per ton of wheat production in rainfed than irrigated system (the results of Table 4 showed this fact). By dividing the amount of each input into wheat production in each hectare (4.8 for irrigated and 1.8 for rainfed farms), it can be seen that more energy has been consumed for production of one ton rainfed than irrigated wheat. The share of each input on final

environmental pollution based on different farm sizing and the hotspots in each category are shown in Figs 7-8 and Table 7.

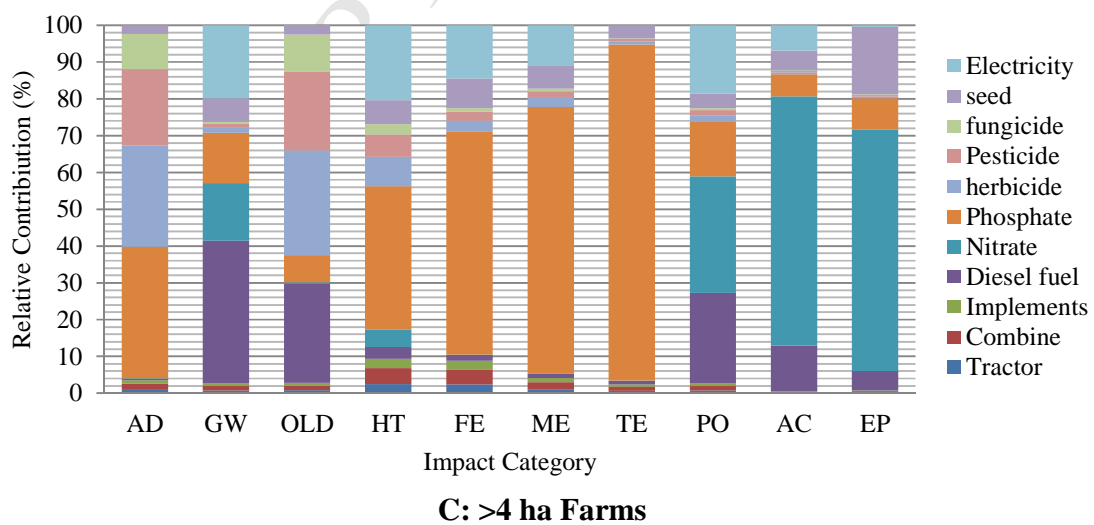
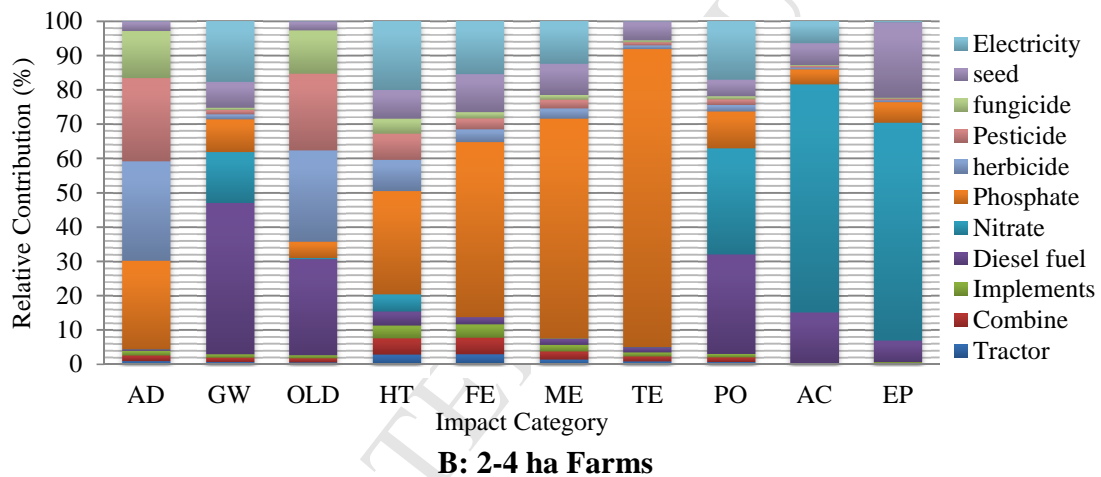
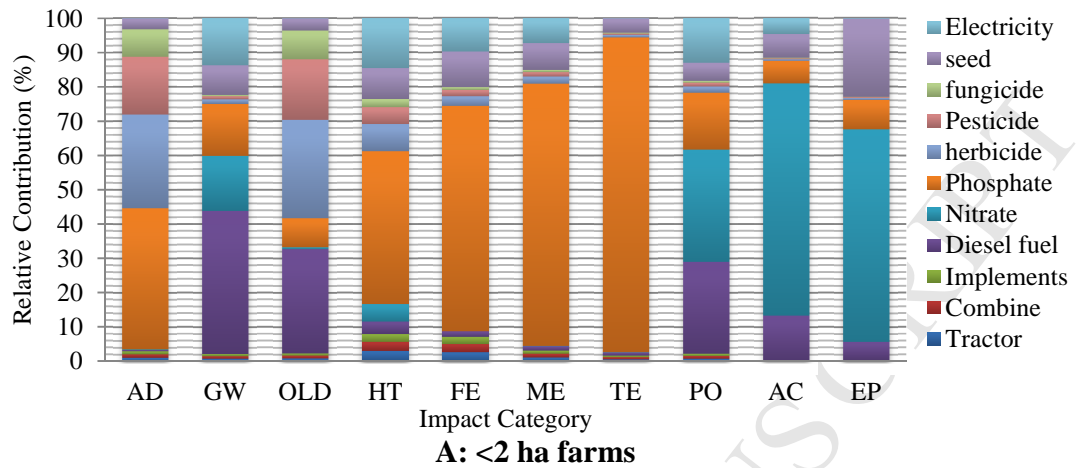


Fig.7. Relative contributions of irrigated wheat production (%) to each impact category (Combine means combine harvester and implements are agricultural machineries expect tractor and combine harvester)

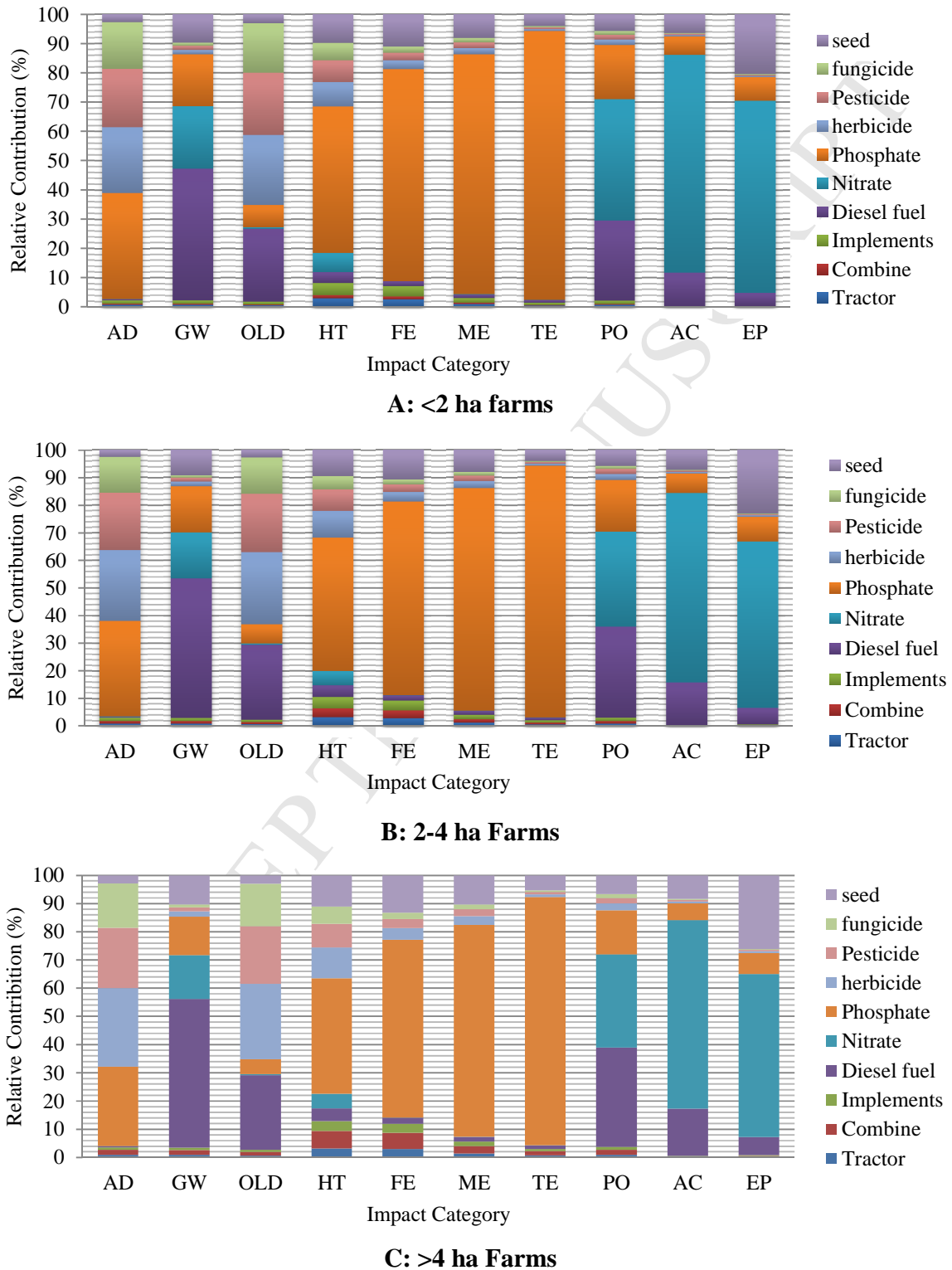


Fig.8. Relative contributions of rainfed wheat production (%) to each impact category (Combine means combine harvester and implements are agricultural machineries expect tractor and combine harvester)

Table.7. Environmental hotspots for irrigated and rainfed wheat production in each impact category.

Impact categories	Irrigated wheat			Rainfed wheat		
	<2 ha	2-4 ha	>4 ha	<2 ha	2-4 ha	>4 ha
AD	Phosphate, Herbicide, Pesticide	Phosphate, Herbicide and Pesticide	Phosphate, Herbicide and Pesticide	Phosphate, Herbicide, Pesticide	Phosphate, Herbicide, Pesticide	Phosphate, Herbicide, Pesticide
GW	Diesel fuel, Nitrate and Phosphate	Diesel fuel and Electricity	Diesel fuel, Nitrate and Electricity	Diesel fuel, Nitrate	Diesel fuel, Nitrate and Phosphate	Diesel fuel, Nitrate and Phosphate
OLD	Diesel fuel, Herbicide and Pesticide	Diesel fuel, Herbicide and Pesticide	Herbicide, Diesel fuel and Pesticide	Diesel fuel, Herbicide and Pesticide	Diesel fuel, Herbicide and Pesticide	Herbicide, Diesel fuel and Pesticide
HT	Phosphate and Electricity	Phosphate and Electricity	Phosphate and Electricity	Phosphate and Seed	Phosphate and Herbicide	Phosphate and Seed
FE	Phosphate, Seed and Electricity	Phosphate, Electricity and Seed	Phosphate and Electricity	Phosphate and Seed	Phosphate and Seed	Phosphate and Seed
ME	Phosphate, seed and Electricity	Phosphate, Electricity and Seed	Phosphate and Electricity	Phosphate and Seed	Phosphate and Seed	Phosphate and Seed
TE	Phosphate and Seed	Phosphate and Seed	Phosphate	Phosphate	Phosphate	Phosphate
PO	Nitrate, Diesel fuel and Phosphate	Nitrate, Diesel fuel and Electricity	Nitrate, Diesel fuel and Electricity	Nitrate, Diesel fuel and	Nitrate, Diesel fuel and	Diesel fuel and Nitrate
AC	Nitrate and Diesel fuel	Nitrate, Diesel fuel and	Nitrate, Diesel fuel and	Nitrate, Diesel fuel and	Nitrate, Diesel fuel and Seed	Nitrate, Diesel fuel and
EP	Nitrate, Seed and Phosphate	Nitrate, Seed	Nitrate, Seed and Phosphate	Nitrate and Seed	Nitrate and Seed	Nitrate and Seed

Toxicity assessment with LCA method is widely used to evaluate agricultural input toxicity. The leachate can increase the potential toxic compounds in groundwater and surface water (Kjeldsen et al., 2002). As it can be seen in Figs. 6 and 7, the most crucial factors in creating the pollution are phosphate and electricity. For this reason, innovative ideas for using the chemical fertilizers and irrigation system in sustainable wheat production should be proposed. The region where wheat production is made is very good in terms of solar energy potential. Solar energy-assisted electricity can be applied in the irrigation as a solution proposal. In terms of agro-chemical pollutions, variable rate technologies and precision application of pesticides and fertilizers can be considered as a potential alternative. Özilgen (2017), noted that microbial fertilizers could replace the chemical fertilizers in the context of sustainable agricultural practices. In this context,

the use of microbial fertilizers instead of chemical fertilizers in the wheat production process could be an innovative application for agricultural applications in this region.

Hoshyar and Grundmann (2017) reported that more than 40 and 70% of total toxicities could be due to ploughing, harrowing, planting and combine harvesting in wheat production by conventional tillage method. The authors reported that planting and combine harvesting applications have higher impacts on the toxicities in no-tillage wheat production system (around 70%). Adom et al. (2012) found that N_2O released from total farm application can cause 65% of GHG emissions, whereas 35% was produced by fertilizer manufacture. As it can be seen in Table 6, GW was obtained as 317.81 and 380.17 kg CO_2 eq. for 1 ton irrigated and rainfed wheat, respectively. This was comparable with the results of other studies (Biswas et al., 2008). According to Nabavi-Pelesaraei et al. (2016), the optimum carbon footprint was identified about 81 kg CO_2 eq. per ton for irrigated wheat. It shows inefficient use of inputs in wheat production in the region.

In this study, the rate of greenhouse gas (GHG) produced by irrigated and rainfed wheat production systems is estimated at 317.81 and 380.17 kg CO_2 eq per 1 ton, respectively. This was comparable with the results of other studies (Biswas et al., 2008). It was found that in this study, diesel fuel followed by nitrate and electricity are the main sources in this index. Some researchers reported that synthetic N fertilizers for cropping is the main factor for calculating the GHG index (Sheehan et al. 1998; Braschkat et al. 2003; Robertson et al, 2000).

The rate of AD impact was found to be 0.002 kg Sb eq. for irrigated and 0.003 kg Sb eq. for rainfed wheat system and the main contribution was due to phosphate followed by electricity and pesticide for both cropping systems. Fantin et al. (2016) calculated AD as 0.0034 kg Sb eq. in Italian wheat production system that is comparable with present results. It can be attributed because of nitrate and phosphate fertilizers used in farm. Based on Hoshyar and Grundmann (2017), nitrate and phosphate are the main factors which affect the AD index on conventional wheat production.

The rate of EP impact was found to be 2.23 and 3.18 kg PO_4^{3-} eq. for irrigated and rainfed wheat production system, respectively. NO_x and NH_3 depositions have the highest impact on terrestrial eutrophication (Potting et al., 2000). Minimizing the phosphorus and nitrogen losses is an effective method to reduce EP. Since the impact of nitrogen on EP was higher than phosphorous, it is suggested that excessive nitrogen consumption should be avoided to achieve better environmental outcome. Hoshyar and Grundmann (2017) reported that farm machinery operations, nitrogen and seed are the key influencing parameters for calculating EP index in wheat production.

Acidification index can explain the climate change. The major compounds of acidification index are SO_2 , NO_x , HCl and NH_3 . (Nabavi-Plesaraei et al, 2017). As indicated in Table 6, the values of AC were 8.991 kg SO_2 eq. for irrigated and 11.863 kg SO_2 eq. for rainfed wheat production system, respectively. The microbial oxidation of fertilizers is the main acid forming reaction for fertilizers (Taki et al, 2016). In this research, more than 60% of this index was due to nitrate and phosphate uses and about 20% is because of diesel fuel and machinery.

Ozone layer depletion is a major problem in climate change and one of the main reasons of global warming. Chlorofluorocarbons (CFCs) have a major impact on ozone layer depletion. (Taki et al, 2012a). In this study, OLD was calculated about 0.00002 and 0.00004 kg CFC11 eq. for irrigated and rainfed wheat production system.

Photochemical oxidation is often a secondary parameter in air pollution. The amounts of PO were calculated 0.145 kg C_2H_4 eq. for irrigated and 0.174 kg C_2H_4 eq. for rainfed wheat

production system, respectively. Nitrate and diesel fuel had the highest impact on PO in both of irrigated and rainfed wheat production. Decreasing the amount of fertilizer and using of farm machinery operation also well pumps can affect this factor.

A general comparison is carried out to compare results of the environmental assessment of wheat production in Iran and other countries using LCA method. Meisterling et al. (2009) investigated the conventional and organic wheat production systems in the USA. The authors reported that, considering global warming, production of 1 kg bread under organic system, can produce only 30 kg CO₂ eq., which is less than the conventional cropping system. Wang et al. (2015) analyzed the GWP of wheat production in north China. GWP value per hectare was reported 2.99-4.59 ton CO₂ eq ha⁻¹.

Overall comparison of two wheat production systems (rainfed and irrigated) in this research showed that rainfed system may cause more environmental burdens than irrigated one and using irrigated system can reduce environmental impacts about 24% compared with rainfed. Results showed that the size of farms has no significant effect on environmental pollutions in both systems. In irrigated system, small farms, created more environmental pollution than other sizes (about 4%), but in rainfed system, environmental burdens were almost same in all three farm size categories.

The overall results of this study indicated that rainfed wheat could produce more environmental pollution than irrigated wheat production system.

Considering the quality of inputs can lead to different results from the findings of this study. One of the methods that can use to optimize the energy and environmental pollutants in terms of the quality of inputs, is thermodynamic approach (Yildizhan, 2017a,b). As a further research, it is recommended to used exergy approach for wheat and other agricultural products for calculation the impact of production on environmental pollution according to the quality of each input. Also a detailed analysis of agro-chemicals (fertilizers and pesticides) is necessary in terms of emissions to each compartment such as soil, water and crop.

4. Conclusion

This paper presented a LCA study of irrigated and rainfed wheat production systems in center of Iran, Mahyar plain, throughout the first phases of their life cycle. The study was used a comprehensive data collection include 210 irrigated and 90 rainfed wheat farms. Based on the results, it has been found that the total input energy for irrigated and rainfed wheat production were 23.41 and 9.35 GJ/ha, respectively. Diesel fuel and total chemical fertilizer were the major inputs in all irrigated and rainfed farm lands. Results showed that energy productivity in irrigated and rainfed wheat production were 132.63 and 133.50 kg GJ⁻¹, respectively. Also the results showed that the share of renewable energy in irrigated was higher than rainfed system. In all categories, the average of renewable energy in irrigated and rainfed wheat production is 4.80 and 1.16 GJ/ha, respectively. The results of LCA method indicated that the rate of greenhouse gas (GHG) produced by irrigated and rainfed wheat production is estimated at 317.81 and 380.17 kg CO₂ eq per 1 ton of produced wheat, respectively. LCA showed that based on 1 ton of wheat production as FU, rainfed system could actually cause more environmental burdens compared to irrigated system. The most crucial factors in creating this pollution are phosphate and electricity. For this reason, innovative ideas for using of chemical fertilizers and irrigation water in sustainable wheat production should be proposed. As a solution for decreasing the GHG emission, incentive policies for the use of renewable energy in agriculture must be presented by decision-makers. It is recommended that farmers be educated about the use of chemical

fertilizers. As another recommendation, the use of microbial fertilizers should be widespread rather than the use of chemical fertilizers during the crop production process. Finally, for sustainable agriculture, farmers need to change their irrigation systems and increase harvesting efficiency to reduce energy consumption.

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References

1. Abdi, R., Taki, M., Akbarpour, M., 2012. An Analysis of Energy input-output and Emissions of Greenhouse Gases from Agricultural Productions. *Int. J. Nat. Eng. Sci.* 6(3): 73-79.
2. Abdi, R., Taki, M., Jalali, A., 2013. Study on energy use pattern, optimization of energy consumption and CO₂ emission for greenhouse tomato production. *Int. J. Nat. Eng. Sci.* 7(1): 01-04.
3. Abdollahpour S., Zaree S., 2009. Evaluation of wheat energy balance under rain fed farming in Kermanshah. *Journal of Sustainable Agriculture Knowledge*; 20(2): 97-106 [in parsian].
4. Achten, W.M.J., Van Acker, K., 2016. EU-Average Impacts of Wheat Production. A Meta-Analysis of Life Cycle Assessments. *J. Ind. Ecol.* 20(1): 132–144.
5. Adom, F., Maes, A., Workman, C., Clayton-Nierderman, Z., Thoma, G., Shonnard, D.R., 2012. Regional carbon footprint analysis of dairy feeds for milk production in the USA. *Int. J. Life Cycle Assess.* 17(5): 520-534.
6. Asgharipour, M.R., Mondani, F., Riahinia, S., 2012. Energy use efficiency and economic analysis of sugar beet production system in Iran: A case study in Khorasan Razavi province. *Energy* 44: 1078–1084.
7. Biswas, W., Barton, L., Carter, D., 2008. Global Warming Potential of Wheat Production in Western Australia: A Life Cycle Assessment. *Water Environ. J.* 22: 206-216.
8. Blanco, J., Inaba, A., Finkbeiner, M., 2015. Scoping organizational LCA—Challenges and solutions, *Int. J. Life Cycle Assess.* 20: 829–841.
9. Braschkat, J., Braschkat, A., Quirin, M., Reinhardt, G.A., 2003. Life cycle assessment of bread production—a comparison of eight different scenarios. Life cycle assessment in the agric-food sector. In Proceedings from the 4th International Conference, 6–8 October, Denmark.
10. Brentrup, F., Kusters, J., Kulmann, H., Lamme, J., 2004a. Environmental impact assessment of agricultural production systems using the life cycle assessment methodology: I. Theoretical concept of a LCA method tailored to crop production. *Eur. J. Agron.* 20(3): 247–264.
11. Brentrup, F., Küsters, J., Lammel, J., Barraclough, P., Kuhlmann, H., 2004b. Environmental impact assessment of agricultural production systems using the life cycle assessment (LCA) methodology II. The application to N fertilizer use in winter wheat production systems. *Eur. J. Agron.* 20, 265-279.
12. Bundschuh, J., Chen, G., (book editors), 2014. Sustainable Energy Solutions in Agriculture, CRC Press, Taylor & Francis Books.
13. Charles, R., Jolliet, O., Gaillard, G., Pellet, D., 2006. Environmental analysis of intensity level in wheat crop production using life cycle assessment. *Agric. Ecosyst. Environ.* 113: 216-225. <http://dx.doi.org/10.1016/j.agee.2005.09.014>
14. Chen, G., Maraseni, T., Yang, Z., 2010. Life-Cycle Energy and Carbon Footprint Assessments: Agricultural and Food Products. In: Capehart, B. (Editor). *Encyclopedia of Energy Engineering and Technology*, 1:1,1-5, Taylor & Francis Books, London, UK.
15. Cleary, J., 2009. Life cycle assessments of municipal solid waste management systems: a comparative analysis of selected peer-reviewed literature. *Environ. Int.* 35(8): 1256-1266.
16. Durlinger, B., Tyszler, M., Scholten, J., Broekema, R., Blonk, H., Beatrixstraat, G., 2014. Agri-Footprint; a Life Cycle Inventory database covering food and feed production and processing. In Proceedings of the 9th International Conference on Life Cycle Assessment in the Agri-Food Sector.

17. European Commission, 2010. International reference life cycle data system (ILCD) handbook. General guide for life cycle assessment—detailed guidance. 1st ed. European Commission–Joint Research Centre–Institute for Environment and Sustainability, Italy.
18. Fantin, V., Righi, S., Rondini, I., Masoni, P., 2016. Environmental assessment of wheat and maize production in an Italian farmers' cooperative. *J. Cleaner Prod.* 140: 631–643. doi: 10.1016/j.jclepro.2016.06.136.
19. Ghorbani, R., Mondani, F., Amirmoradi, S., Feizi, H., Khorramdel, H., Teimouri, S., Sanjani, M., Anvarkhah, S., Aghel, S., 2011. A case study of energy use and economical analysis of irrigated and dryland wheat production systems. *Appl. Energy* 88: 283–288.
20. Guinee, J.B., 2002. Handbook on life cycle assessment operational guide to the ISO standards, *Int. J. Life Cycle Assess.* 7: 311–313.
21. Hoshyar, E., Grundmann, P., 2017. Environmental Impacts of Energy Use in Wheat Tillage Systems: A Comparative Life Cycle Assessment (LCA) Study in Iran. *Energy.* 122: 11–24. doi: 10.1016/j.energy.2017.01.069.
22. ISO (International Organization for Standardization), 2006. Environmental management- Life Cycle Assessment- Principles and Framework. ISO 14040.
23. Khoshnevisan B, Rafiee S, Omid M, Mousazadeh H. 2013b. Applying data envelopment analysis approach to improve energy efficiency and reduce GHG (greenhouse gas) emission of wheat production. *Energy* 58: 588–593.
24. Khoshnevisan B, Rafiee S, Omid M, Yousefi M, Movahedi M. 2013c. Modeling of energy consumption and GHG (greenhouse gas) emissions in wheat production in Isfahan province of Iran using artificial neural networks. *Energy* 2013: 333–338.
25. Khoshnevisan, B., Rafiee, S., Omid, M., Mousazadeh, H., Shamsheerband, S., Hamid, S.H., 2015. Developing a fuzzy clustering model for better energy use in farm management systems. *Renew. Sustain. Energy Rev.* 48: 27–34.
26. Khoshnevisan, B., Rafiee, Sh., Omid, M., Yousefi, M., 2013. Prediction of environmental indices of Iran wheat production using artificial neural networks. *Int. J. Energy Environ.* 4(2): 339–48.
27. Kim, S., Dale, B.E., Jenkins, R., 2009. Life cycle assessment of corn grain and corn stover in the United States. *Int. J. Life Cycle Assess.* 14: 160–174.
28. Kjeldsen, P., Barlaz, M.A., Rooker, A.P., Baun, A., Ledin, A., Christensen, T.H., 2002. Present and long-term composition of MSW landfill leachate: a review. *Crit. Rev. Environ. Sci. Tec.* 32(4): 297–336.
29. Koocheki, A., Ghorbani, R., Mondani, F., Alizade, Y., Moradi, M., 2011. Pulses production systems in term of energy use efficiency and economical analysis in Iran. *Int. J. Energy Econ. Policy.* 1: 95–106.
30. Kouchaki-Penchah, H., Nabavi-Pelesaraei, A., O'Dwyer, J., Sharifi, M., 2017. Environmental management of tea production using joint of life cycle assessment and data envelopment analysis approaches. *Environ. Prog. Sustainable Energy.* 36(4): 1116–1122. DOI 10.1002/ep.12550.
31. Kuesters, J., Lammel, J., 1999. Investigations of the energy efficiency of the production of winter wheat and sugar beet in Europe. *Eur. J. Agron.* 11: 35–43.
32. Liamsanguan, C., Gheewala, S.H., 2008. LCA: A decision support tool for environmental assessment of MSW management systems. *J. Environ. Manage.* 87(1): 132–138.
33. Maraseni, T.N., Chen, G., Banhazi, T., Bundschuh, J., Yusaf, T., 2015. An assessment of direct on-farm energy use for high value grain crops grown under different farming practices in Australia. *Energies.* 8: 13033–13046.
34. Meisterling, K., Samaras, C., Schweizer, V., 2009. Decisions to reduce greenhouse gases from agriculture and product transport: LCA case study of organic and conventional wheat. *J. Cleaner Prod.* 17: 222–230.
35. Mohammadi, A., Rafiee, S., Jafari, A., Keyhani, A., Mousavi-Avval, S.H., Nonhebel, S., 2014. Energy use efficiency and greenhouse gas emissions of farming systems in north Iran. *Renew. Sustain. Energy Rev.* 30: 724–733.

36. Mondany, F., Aleagha S., Khoramivafa, M., Ghobadi, R., 2017. Evaluation of greenhouse gases emission based on energy consumption in wheat Agroecosystems. *Energy Reports* 3: 37–45.
37. Nabavi-Pelesaraei, A., Abdi, R., Rafiee, S., Mobtaker, H., 2013. Optimization of energy required and greenhouse gas emissions analysis for orange producers using data envelopment analysis approach. *J. Cleaner Prod.* 65: 311-317. <http://dx.doi.org/10.1016/j.jclepro.2013.08.019>.
38. Nabavi-Pelesaraei, A., Bayat, R., Hosseinzadeh-Bandbafha, H., Afrasyabi, H., Chau, K.W., 2017. Modeling of energy consumption and environmental life cycle assessment for incineration and landfill systems of municipal solid waste management - A case study in Tehran Metropolis of Iran. *J. Cleaner Prod.* 148: 427-440.
39. Nabavi-Pelesaraei, A., Hosseinzadeh-Bandbafha, H., Qasemi-Kordkheili, P., Kouchaki-Penchah, H., Riahi-Dorcheh, F., 2016. Applying optimization techniques to improve of energy efficiency and GHG (greenhouse gas) emissions of wheat production. *Energy.* 103: 672-678.
40. Nabavi-Pelesaraei, A., Kouchaki-Penchah, H., Amid, S., 2014. Modeling and optimization of CO₂ emissions for tangerine production using artificial neural networks and data envelopment analysis. *International Journal of Biosciences.* 4: 148–158.
41. Nawi, N.M, Yahya, A., Chen, G., Bockari-Geva, S.M., Maraseni, T.N., 2012. Human Energy Expenditure in Lowland Rice Cultivation in Malaysia. *J. Agric. Saf. Health.* 18(1): 45-56.
42. Nemecek, T., Huguenin-Elie, O., Dubois, D., Gaillard, G., Schaller, B., Chervet, A., 2011. Life cycle assessment of Swiss farming systems: II. Extensive and intensive production. *Agric Syst* 104(3): 233-45.
43. Nemecek, T., Kagi, T., Blaser, S., 2007. Life cycle inventories of agricultural production systems. Final Report Ecoinvent V2.0 No. 15. Agroscope Reckenholz-Taenikon Research Station ART, Swiss Centre for Life Cycle Inventories, Zurich and Dübendorf, CH.
44. Nielsen, P.H., Weidema, B.P., Nielsen, A.M., Dalgaard, R., Halberg, N., 2005. LCA food database. (Accessed January 2016), <http://www.lcafood.dk>.
45. Özilgen, M., 2017. Assesment of the use of zero-emission vehicles and microbial fertilizers in beverage produvtion. *J. Cleaner Prod.* 165: 298-311.
46. Pahlavan R., Omid, M., Rafiee, S., Mousavi-Avval, S.H., 2012. Optimization of energy consumption for rose production in Iran. *Energy Sustain Dev.* 16(2): 236-41.
47. Pahlavan, R., Omid, M., Akram, A., 2011. Energy use efficiency in greenhouse tomato production in Iran. *Energy.* 36: 6714–6719.
48. Ramedani, Z., Rafiee, S., Heidari, M.D., 2011. An investigation on energy consumption and sensitivity analysis of soybean production farms. *Energy.* 36(11): 6340-4.
49. Ranjbar, I., Ajabshirchi, Y., Taki, M., Ghobadifar, A., 2013. Energy consumption and modeling of output energy with MLP Neural Network for dry wheat production in Iran. *Elixir Agriculture* 62: 17949-17953.
50. Rebitzer, G., Ekvall, T., Frischknecht, R., Hunkeler, D., Norris, G., Rydberg, T., Pennington, D.W., 2004. Life cycle assessment: Part 1: Framework, goal and scope definition, inventory analysis, and applications. *Environ. Int.* 30(5): 701-720.
51. Rezvani-Moghaddam, P., Feizi, H., Mondani, F., 2011. Evaluation of tomato production systems in terms of energy use efficiency and economical analysis in Iran. *Not. Sci. Biol.* 3: 58–65.
52. Rivela, B., Moreira, M.T., Munoz, I., Rieradevall, J., Feijoo, G., 2006. Life cycle assessment of wood wastes: A case study of ephemeral architecture. *Sci. Total Environ.* 357: 1–11.
53. Robertson, G.P., Paul, E.A., Harwood, R.R., 2000. Greenhouse gases in intensive agriculture: Contributions of individual gases to the radiative forcing of the atmosphere. *Science.* 289: 1922–1925.
54. Safa, M., Samarasinghe, S., 2011. Determination and modelling of energy consumption in wheat production using neural networks: “A case study in Canterbury province, New Zealand”. *Energy.* 36(8): 5140-5147.
55. Safa, M., Samarasinghe, S., Mohsen, M., 2009. A field study of energy consumption in wheat production in Canterbury, New Zealand. *Energy Convers. Manage.* 52: 2526–2532.

56. Sheehan, J., Camobreco, V., Duffield, J., Graboski, M., Shapouri, H., 1998. Life cycle inventory of biodiesel and petroleum diesel for use in an urban bus. Colorado: NREL/SR-580-24089 UC Category 1503, National Renewable Energy Laboratory.
57. Taghavifar, H., Mardani, A., 2015. Energy consumption analysis of wheat production in West Azarbayjan utilizing life cycle assessment (LCA). *Renewable Energy* 74: 208-213.
58. Taki M, Rohani A, Soheili-Fard F, Abdeshahi A. 2018. Assessment of energy consumption and modeling of output energy for wheat production by neural network (MLP and RBF) and Gaussian process regression (GPR) models. *Journal of cleaner production* 172: 3028-3041.
59. Taki, M., Abdi, R., Akbarpour, M., Mobtaker, H.G., 2013. Energy inputs – yield relationship and sensitivity analysis for tomato greenhouse production in Iran. *Agric Eng Int: CIGR Journal* 15(1): 59-67
60. Taki, M., Ajabshirchi, Y., Ghobadifar, A., 2016. Application of nonparametric method for optimization of energy consumption and greenhouse gas emission in wheat production. *J. Environ. Sci. Technol.* 18(2): 101-114.
61. Taki, M., Ajabshirchi, Y., Mahmoudi, A., 2012a. Prediction of output energy for wheat production using artificial neural networks in Isfahan province of Iran. *J. Agri. Tech.* 8(4): 1229-1242.
62. Taki, M., Ajabshirchi, Y., Mahmoudi, A., 2012b. Application of Parametric and Non-parametric Method to Analyzing of Energy Consumption for cucumber Production in Iran. *Mod. Appl. Sci.* 6(1): 75-87.
63. Taki, M., Mahmoudi, A., Ghasemi-mobtaker, H., Rahbari, H., 2012c. Energy consumption and modeling of output energy with multilayer feed-forward neural network for corn silage in Iran. *Agric Eng Int: CIGR Journal.* 14(4): 93-101.
64. Wang, Z., Zhang, H., Lu, X., Wang, M., Chu, Q., Wen, X., Chen, F., 2015. Lowering carbon footprint of winter wheat by improving management practices in North China Plain. *J. Clean. Prod.* 112: 149-157. <http://dx.doi.org/10.1016/j.jclepro.2015.06.084>.
65. Wernet, G., Bauer, C., Steubing, B., Reinhard, J., Moreno-Ruiz, E., Weidema, B., 2016. The ecoinvent database version 3 (part I): overview and methodology. *Int. J. Life Cycle Assess.* 21(9): 1218-1230.
66. Yildizhan, H., 2017a. Energy, Exergy Utilization and CO₂ Emission of Strawberry Production in Greenhouse and Open Field. *Energy.* 143: 417-423. doi: 10.1016/j.energy.2017.10.139
67. Yildizhan, H., 2017b. Thermodynamics Analysis For A New Approach to Agricultural Practices: Case of Potato Production. *J. Cleaner Prod.* 166: 660-667. doi:10.1016/j.jclepro.2017.08.082
68. Yuan, S., Peng, S., 2017. Input-output energy analysis of rice production in different crop management practices in central China. *Energy.* 141: 1124-1132. <https://doi.org/10.1016/j.energy.2017.10.007>.